

Probing into the impact of 3DVAR assimilation of surface PM₁₀ observations over China using process analysis

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[1] The capability of assimilating surface PM₁₀ (particulate matter with diameters less than 10 μm) observations has been developed within the National Centers for Environmental Prediction Gridpoint Statistical Interpolation three-dimensional variational (3DVAR) data assimilation (DA) system. It provides aerosol analyses for the Goddard Chemistry Aerosol Radiation and Transport aerosol scheme within the Weather Research and Forecasting/Chemistry model. Control and assimilation experiments were performed for June 2011 over China to explore in detail the impact of assimilating surface PM₁₀. In the assimilation experiment, analyses were produced every 6 h to adjust the mass concentrations of different aerosol species. The statistical results from two parallel experiments demonstrate that the assimilation of surface PM₁₀ observations can significantly reduce the uncertainty of initial aerosol fields and effectively improve the subsequent aerosol forecasts for at least 12 h. However, the benefit from the assimilation of PM₁₀ diminishes rapidly with forecast range. Process analysis for PM₁₀ formation indicates that the rapidly diminishing DA impact on aerosol forecasts, especially in early forecast hours, was dominated by vertical mixing with an additional contribution from advection. Both processes were mainly related to unbalanced aerosol fields in the horizontal and vertical after assimilating surface observations. These findings illustrate the importance of adjusting aerosol emission rates and the initial aerosol vertical profile.

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1. Introduction

[2] Considerable progress has been made in recent years to reduce large uncertainties involved with numerical prediction of atmospheric aerosols [e.g., *Hakami et al.*, 2005; *Henze et al.*, 2007, 2009; *Yumimoto et al.*, 2007, 2008a, 2008b; *Dubovik et al.*, 2008; *Benedetti et al.*, 2009; *Pagowski et al.*, 2010; *Pagowski and Grell*, 2012; *Liu et al.*, 2011; *Schwartz et al.*, 2012]. These uncertainties can usually be attributed to inaccurate aerosol emissions and initial conditions (ICs) and deficiencies in the modeling system (e.g., nonlinear physical processes such as advection, diffusion, radiative effects, and cloud and precipitation formation). Data assimilation (DA), as a method to improve the model ICs, has been widely applied to operations and research in numerical weather prediction [e.g., *Parrish and*

Derber, 1992; *Lorenc et al.*, 2000; *Rabier et al.*, 2000; *Gauthier et al.*, 2007; *Kalnay*, 2010] for decades.

[3] However, aerosol DA remains in its infancy. Similar to meteorological DA, a variety of algorithms such as optimal interpolation (OI), two-dimensional variational (2DVAR), three-dimensional variational (3DVAR), four-dimensional variational (4DVAR), and ensemble Kalman filter (EnKF) DA techniques have been used in aerosol DA. Several attempts have recently been made to assimilate aerosol observations into numerical models with the purpose to improve the ICs. However, most of them have focused on assimilating satellite-derived aerosol products, due to their global coverage, especially aerosol optical depth (AOD) (e.g., OI [Collins et al., 2001; Yu et al., 2003; Generoso et al., 2007; Adhikary et al., 2008], Newtonian-nudging [Wang et al., 2004], 2DVAR [Zhang et al., 2008; Schroedter-Homscheidt et al., 2010], 3DVAR [Liu et al., 2011; Schwartz et al., 2012], and 4DVAR [Benedetti et al., 2009]). For instance, Liu et al. [2011] developed a 3DVAR algorithm to assimilate AOD retrievals from Moderate Resolution Imaging Spectroradiometer (MODIS) sensors while studying a dust storm in East Asia.

[4] Comparatively less work has been done to assimilate surface aerosol observations into ICs. Lin et al. [2008a] assimilated surface PM₁₀ (particulate matter with diameters

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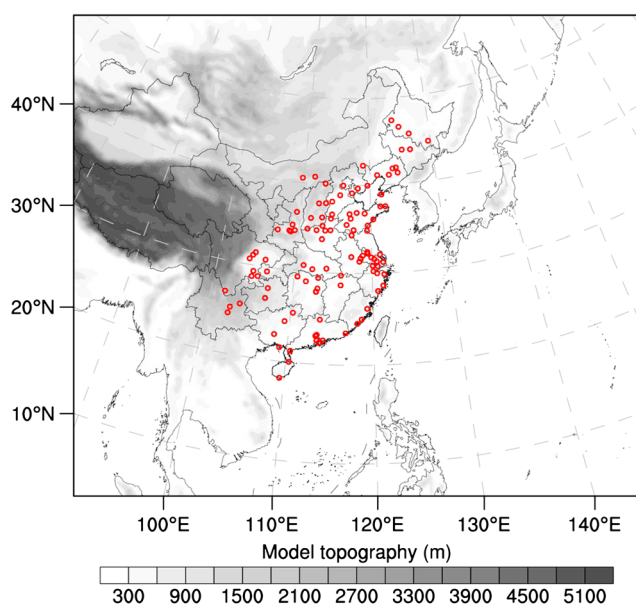


Figure 1. The model domain and the observation network with model topography (m). The open circles depict locations of measurement sites used for PM₁₀ assimilation cycles and PM₁₀ forecast verification.

less than 10 μm) observations in North China with an EnKF. Using the OI method, Tombette *et al.* [2009] assimilated surface PM₁₀ over Europe and Lee *et al.* [2013] assimilated PM₁₀ over South Korea. Pagowski *et al.* [2010] applied a 3DVAR approach to assimilate surface PM_{2.5} (particulate matter with diameters less than 2.5 μm) observations over the continental United States (CONUS), and Pagowski and Grell [2012] used an EnKF for PM_{2.5} DA over the CONUS. Moreover, both MODIS AOD and surface PM_{2.5} observations were assimilated separately and together over the CONUS to illustrate the synergistic effect of assimilating different aerosol observations on aerosol forecasts [Schwartz *et al.*, 2012]. These studies mostly demonstrated that along with successful DA of aerosol-related observations from both ground networks and satellite platforms, the adjusted aerosol ICs substantially improved subsequent aerosol forecasts at short ranges (~ 1 –2 day) but not at extended periods. Also, when solely assimilating surface particulate matter observations, the DA impact on aerosol forecasts fades very quickly in the early forecast hours [Tombette *et al.*, 2009; Pagowski *et al.*, 2010; Schwartz *et al.*, 2012; Lee *et al.*, 2013].

[5] However, the studies mentioned above did not provide quantitative explanations regarding the rapidly diminishing DA impact of surface aerosol observations. Process analysis (PA) [e.g., Jeffries and Tonnesen, 1994; Jang *et al.*, 1995; Tonnesen, 1995; Tonnesen and Dennis, 2000a, 2000b; Jiang *et al.*, 2003; Huang *et al.*, 2005; Zhang *et al.*, 2009b; Liu *et al.*, 2010], a mass balance analysis technique, may identify the main sources contributing to the rapidly diminishing DA impact by providing quantitative information on the formation mechanisms of gaseous and PM pollutants from various chemical and physical processes. Fewer PA studies have been undertaken that examine the formation of PM pollutants compared to those of gaseous pollutants. Previous studies [e.g., Zhang *et al.*, 2009b; Liu *et al.*, 2010]

indicated that emissions, horizontal transport (including horizontal advection and diffusion), aerosol processes (e.g., gas-to-particle conversion processes), and cloud processes contribute the most to PM production and removal in the atmospheric boundary layer. However, at most surface sites, emission processes and vertical transport (including vertical advection and diffusion) are the predominant contributors to PM accumulation and loss. The contribution from dry deposition is relatively small.

[6] An aerosol 3DVAR DA framework has been developed by Liu *et al.* [2011] (hereafter L11) within the National Centers for Environmental Prediction (NCEP) Gridpoint Statistical Interpolation (GSI) 3DVAR DA system [Wu *et al.*, 2002; Kleist *et al.*, 2009], coupled to the Goddard Chemistry Aerosol Radiation and Transport (GOCART) aerosol scheme [Chin *et al.*, 2000, 2002; Ginoux *et al.*, 2001] within the Weather Research and Forecasting/Chemistry (WRF/Chem) model [Grell *et al.*, 2005]. L11 first implemented MODIS AOD DA and applied it to a dust storm event over East Asia. Schwartz *et al.* [2012] (hereafter S12) further extended the system to allow the simultaneous assimilation of both MODIS AOD and surface PM_{2.5}. In this study, surface PM₁₀ assimilation capability is implemented within the same framework. To the author's knowledge, this is the first attempt to assimilate the hourly ground-based PM₁₀ observations within a 3DVAR DA system. Furthermore, PA capability for PM₁₀ is also added into WRF/Chem to better understand the mechanisms of the fast-fading DA impact on aerosol forecasts.

[7] The next section provides a brief description of the WRF/Chem model and GSI DA system. The technical implementation of PM₁₀ DA in the GSI 3DVAR system is given in section 3. The experimental design is described in section 4. The impact of PM₁₀ DA and PA is detailed in section 5 before concluding in section 6.

2. Modeling and DA Systems

[8] The modeling and DA systems used here were described by L11 and S12. Therefore, generally brief descriptions follow, and important differences are noted.

2.1. WRF/Chem Model Configurations

[9] In this study, version 3.3.1 of the WRF/Chem model [Grell *et al.*, 2005] was used to simultaneously predict weather and atmospheric composition. WRF/Chem is an “online” model, as its chemical and meteorological components are fully coupled. Same as L11 and S12, GOCART was chosen as the aerosol option. The original GOCART simulates 14 tropospheric aerosol types including sulfate,

Table 1. Statistics Comparing the Lowest Model Level PM₁₀ Mass Concentrations From the Control and Assimilation Experiments, Calculated Against Observations From the MEP Network Over All 0000 and 1200 UTC Analyses During 01 to 28 June 2011

Exp. Name	Analyses Time (UTC)	Mean Obs. ($\mu\text{g}/\text{m}^3$)	Mean Sim. ($\mu\text{g}/\text{m}^3$)	BIAS ($\mu\text{g}/\text{m}^3$)	RMSE ($\mu\text{g}/\text{m}^3$)	CORR
Control	0000	77.16	54.66	−22.50	52.94	0.414
	1200	68.88	83.68	14.80	54.70	0.395
Assimilation	0000	77.16	69.31	−7.84	21.32	0.944
	1200	68.88	71.30	2.43	21.19	0.887

organic carbon (OC), black carbon (BC), mineral dust in five particle-size bins (effective radii of 0.5, 1.4, 2.4, 4.5, and 8.0 μm), and sea salt in four particle-size bins (effective radii of 0.3, 1.0, 3.25, and 7.5 μm for dry air). For OC and BC, hydrophobic and hydrophilic components are considered. However, for the GOCART module built in WRF/Chem, two additional variables “p25” and “p10” are also included to account for fine (effective diameter < 2.5 μm) and coarse (effective diameter 2.5–10 μm) mode unspiciated aerosols. Model-output PM₁₀ is diagnosed from 14 of 16 prognostic aerosol variables (see section 3).

[10] The model configurations mostly follow L11 for the model domain (Figure 1), grid spacing (27 km), horizontal grid points (261 $\times$ 222), vertical levels (45, with the model top at 50 hPa), physical parameterizations, and chemistry suite. Aerosol direct effects are allowed through the coupling between GOCART and the Goddard shortwave radiation scheme [Fast *et al.*, 2006; Barnard *et al.*, 2010]. Anthropogenic emissions were supplied offline from an Asia emissions inventory [Streets *et al.*, 2003; Zhang *et al.*, 2009a], and biogenic emissions were calculated online from the U.S. Geological Survey land use classification by using the Guenther scheme [Guenther *et al.*, 1994; Simpson *et al.*, 1995]. In addition, dust emissions [Ginoux *et al.*, 2001], dimethylsulfide, and sea salt emissions [Chin *et al.*, 2000, 2002] were also computed online. Lateral boundary conditions (LBCs) and ICs for meteorological fields were provided by the NCEP Global Forecast System (GFS). Aerosol and chemical ICs originated from previous WRF/Chem forecasts and chemical and aerosol LBCs are based upon an idealized, northern hemispheric, midlatitude, clean environmental vertical profile from the NOAA Aeronomy Lab Regional Oxidant Model [McKeen *et al.*, 1991; Liu *et al.*, 1996].

2.2. Incorporation of PA Within WRF/Chem

[11] To determine the roles of individual physical and chemical processes in species formation, PA has been widely applied to separate rates of change of species’ concentrations into various contribution terms. The WRF/Chem model adopts the K-theory form of the scalar conservation equation to describe the tendency due to each process, given by

$$\frac{\partial C_i}{\partial t} = - \left[u \frac{\partial C_i}{\partial x} + v \frac{\partial C_i}{\partial y} + w \frac{\partial C_i}{\partial z} \right] + \frac{\partial}{\partial z} \left(K_e \frac{\partial C_i}{\partial z} \right) + D + R + E, \quad (1)$$

where the term on the left side is the concentration (C) tendency of species i with respect to time; the first term on the right-hand side represents horizontal and vertical advections in the x , y , and z directions, and u , v , and w are the mean wind speeds in the three directions, respectively; the second term denotes the vertical turbulent diffusion process (the horizontal diffusion terms are neglected), and K_e is turbulent diffusivities; the latter three terms are dry deposition (D), net chemical reactions (R), and source emission rate (E), respectively.

[12] We incorporated the PA capability within WRF/Chem by modifying the model source code to diagnose the tendency of each species concentration due to each process every time step by differencing species’ concentrations before and after each process. The accumulated tendency, integrated

during a period of time in each grid cell, is commonly referred to as the integrated process rate (IPR) [Huang *et al.*, 2005; Zhang *et al.*, 2009b; Liu *et al.*, 2010]. In this paper, the PM₁₀ IPR is examined every hour to explore the relative contributions of major atmospheric processes in PM₁₀ formation. Positive values of IPR indicate PM₁₀ production from various processes, and negative values indicate PM₁₀ losses. Thus, the contribution of each process may be compared directly between parallel experiments, especially for the near-surface layer, to further examine the DA impact on subsequent WRF/Chem aerosol forecasts.

2.3. GSI 3DVAR System

[13] The GSI 3DVAR DA system produces an analysis in model grid space. The analysis is obtained through the minimization of a scalar objective function $J(\mathbf{x})$ given by

$$J(\mathbf{x}) = \frac{1}{2} (\mathbf{x} - \mathbf{x}_b)^T \mathbf{B}^{-1} (\mathbf{x} - \mathbf{x}_b) + \frac{1}{2} [H(\mathbf{x}) - \mathbf{y}]^T \mathbf{R}^{-1} [H(\mathbf{x}) - \mathbf{y}], \quad (2)$$

where $\mathbf{x}$ is the model state and the subscript b denotes the background state, $\mathbf{y}$ is the observation vector, H is the observation operator that computes the observation estimates from the model state vector, $\mathbf{B}$ denotes the background error covariance (BEC) matrix, and $\mathbf{R}$ represents the observation error covariance matrix, including contributions from measurement and representativeness errors (see section 3.2 for details). The error covariance matrices determine the relative contributions of the background and observation terms to the final analysis.

[14] In our case, GSI was used to assimilate surface PM₁₀ observations without meteorological DA. Similar to L11 and S12, the 3-D mass mixing ratios of the 16 GOCART aerosol species at each grid point comprised the analysis (or control) variables in the GSI 3DVAR minimization procedure. This speciated approach to aerosol DA was introduced by L11 and further applied by S12. Different from L11 and S12, an additional variable P₁₀ (coarse-mode unspiciated aerosols) was introduced into the analysis vector here as it is an important contributor to PM₁₀. However, a total of 14 control variables were used to diagnose the model estimate PM₁₀ at the observation locations within the GSI 3DVAR system, as detailed in section 3. Thus, only 14 GOCART aerosol variables were updated in the aerosol DA.

[15] As in L11 and S12, the latitude-dependent BEC statistics for each aerosol species were estimated using the National Meteorological Center (now known as NCEP) method [Parrish and Derber, 1992; Rabier *et al.*, 1998], which takes differences between forecasts of different lengths valid at common times. Utilizing the differences of 24 and 12 h WRF/Chem forecasts of the analysis variables valid at the same time for 59 pairs valid at either 0000 and 1200 UTC over the experimental period (June 2011), standard deviations and horizontal and vertical length scales for each aerosol species’ BEC were calculated. No cross correlation between different aerosol species was considered because of the incapability of the current GSI 3DVAR to directly model the cross correlations in the $\mathbf{B}$ matrix. Work is underway to take into account cross correlation between species through the ensemble-based DA techniques [Pagowski and Grell, 2012].

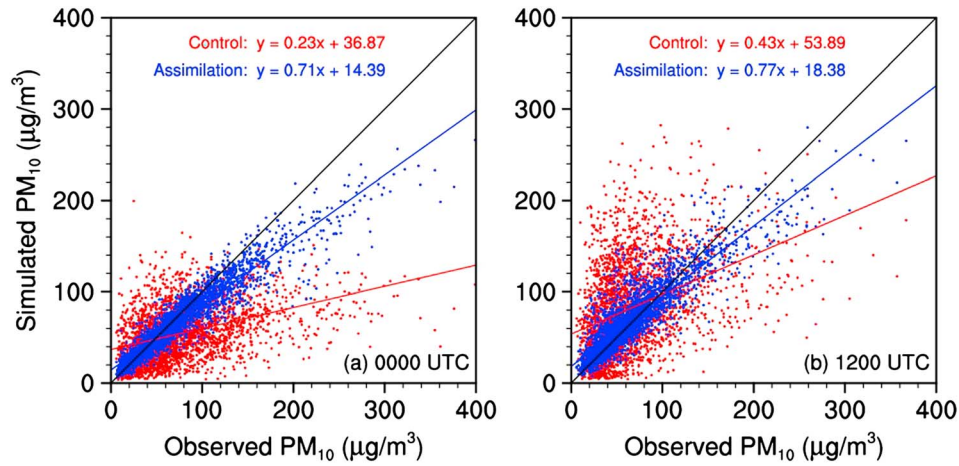


Figure 2. Scatter plots of simulated versus observed PM₁₀ mass concentrations over all (a) 0000 and (b) 1200 UTC initializations from the control (red) and assimilation (blue) experiments.

3. Implementation of PM₁₀ DA

3.1. PM₁₀ Observation Operator

[16] To assimilate PM₁₀, it was necessary to derive model-simulated PM₁₀ at the observation locations. Following the WRF/Chem GOCART aerosol module, the diagnostic variable PM₁₀ is obtained from the 14 of 16 aerosol species, mineral dust and sea salt in the largest particle-size bins were excluded. To assimilate a surface PM₁₀ observation, a model estimate of PM₁₀ is diagnosed at the lowest vertical level by summing mass mixing ratios of aerosol particles, given as

$$PM_{10} = \rho_d[(P_1 + P_2) + (D_1 + D_2 + D_3 + 0.87D_4) + 1.8(O_1 + O_2) + (B_1 + B_2) + (S_1 + S_2 + S_3) + 1.375U], \quad (3)$$

where P_1 and P_2 represent fine- and coarse-mode unspiciated aerosol contributions to PM₁₀, respectively; D_1 , D_2 , and D_3 (S_1 , S_2 , and S_3) are mineral dust (sea salt) aerosols in three smallest particle-size bins, and D_4 denotes mineral dust in the fourth smallest particle-size bin (effective radius of 4.5

µm); O_1 and O_2 (B_1 and B_2) are hydrophobic and hydrophilic OC (BC), respectively; and U denotes sulfate. Similar to S12, coefficients <1 in equation (3) account for the 10 µm diameter cutoff of GOCART aerosols for PM₁₀, and coefficients >1 empirically account for additional fine particulate mass not predicted explicitly by GOCART, such as oxygen contained in organic aerosols (associated with OC) and ammonium (typically associated with sulfate aerosols). The dry air density ρ_d converts the units from µg/kg to µg/m³ for consistency with the observations. Before equation (3) is applied within GSI, the aerosol analysis variables and ρ_d are bilinearly interpolated in the horizontal to the observation locations. Since altitudes of some measurement sites are not available, differences between model topography and reality were neglected, and no vertical extrapolation was performed. Observations taken within 1 h of the analysis were assimilated.

3.2. PM₁₀ Observation Data and Errors

[17] Hourly averaged surface PM₁₀ concentrations for June 2011 were obtained from the Ministry of Environmental Protection (MEP) of China. At each measurement site, the samplings are heated to measure the mass of dry particulate

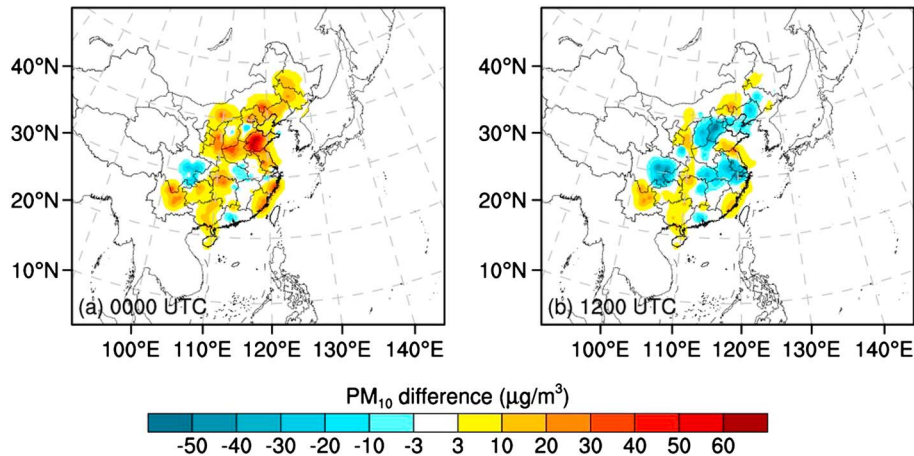


Figure 3. PM₁₀ mass differences (assimilation minus control) at the lowest model level averaged over all (a) 0000 and (b) 1200 UTC initializations.

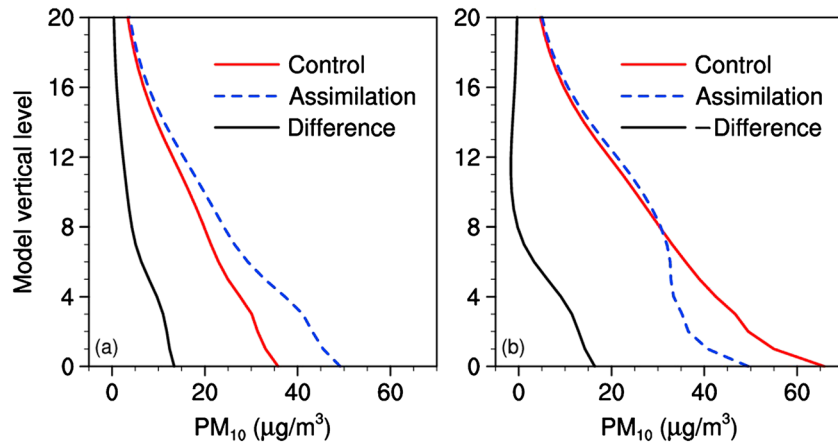


Figure 4. Vertical profiles of simulated PM₁₀ mass concentrations from the control (red) and assimilation (blue) experiments, and the difference (black), averaged over all 0000 and 1200 UTC initializations for locations where there are (a) positive and (b) negative differences at the lowest model level. Note that the legend in Figure 4a differs partly from that of Figure 4b.

matter using the tapered element oscillating microbalance (TEOM) method [Green *et al.*, 2001; Charron *et al.*, 2003]. Figure 1 shows the locations of 112 measurement sites used for the PM₁₀ assimilation experiment and forecast verification. Observation sites span most of central and eastern China and are primarily located in urban and suburban areas.

[18] MEP PM₁₀ data were provided without information regarding data quality. To ensure data reliability before DA,

PM₁₀ observations were subject to quality control (QC). Two QC checks were performed, including value-range and time-continuity checks. First, PM₁₀ observational values that fell outside of subjectively chosen lower ($6 \mu\text{g}/\text{m}^3$) and upper ($425 \mu\text{g}/\text{m}^3$) limits were deemed unrealistic and rejected. Second, time continuity was checked to eliminate gross outliers. Similar to the “buddy check” used in meteorological DA [e.g., Dee *et al.*, 2001], any measurement $O(t)$ will pass

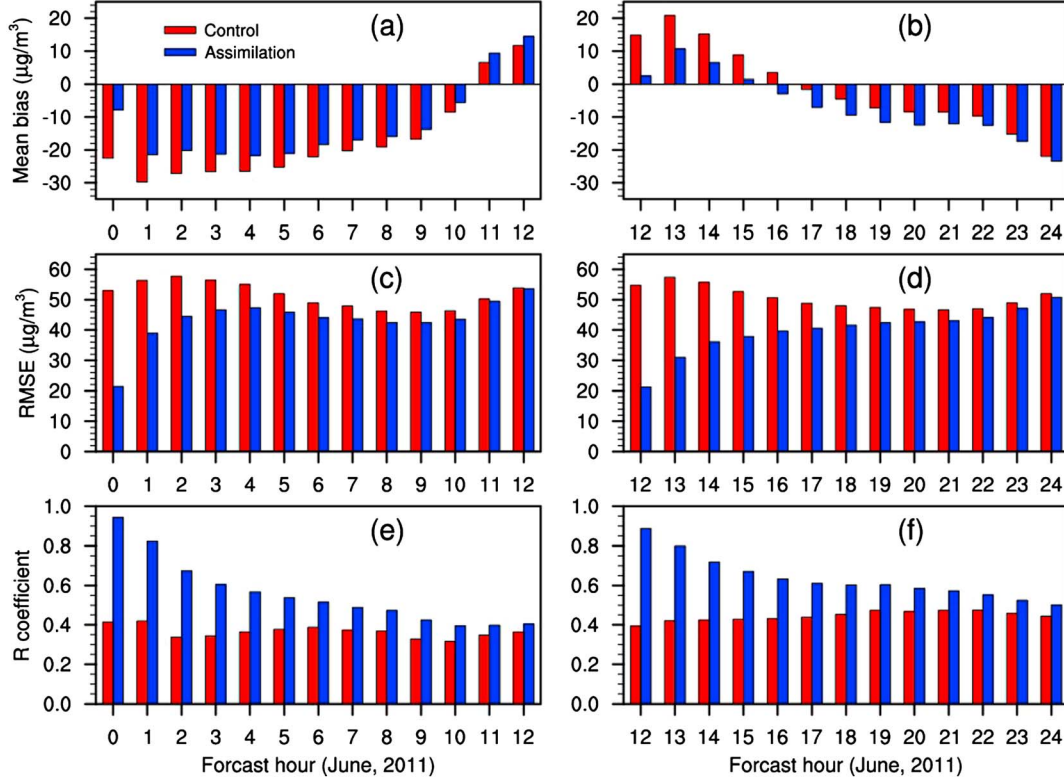


Figure 5. Evaluation statistics of surface PM₁₀ forecasts as a function of forecast range calculated against observations: (top) mean bias ($\mu\text{g}/\text{m}^3$), (middle) root-mean-square error ($\mu\text{g}/\text{m}^3$), and (bottom) correlation coefficient. The forecasts are produced from the 0000 (left) and 1200 (right) UTC initializations during 1 to 28 June 2011.

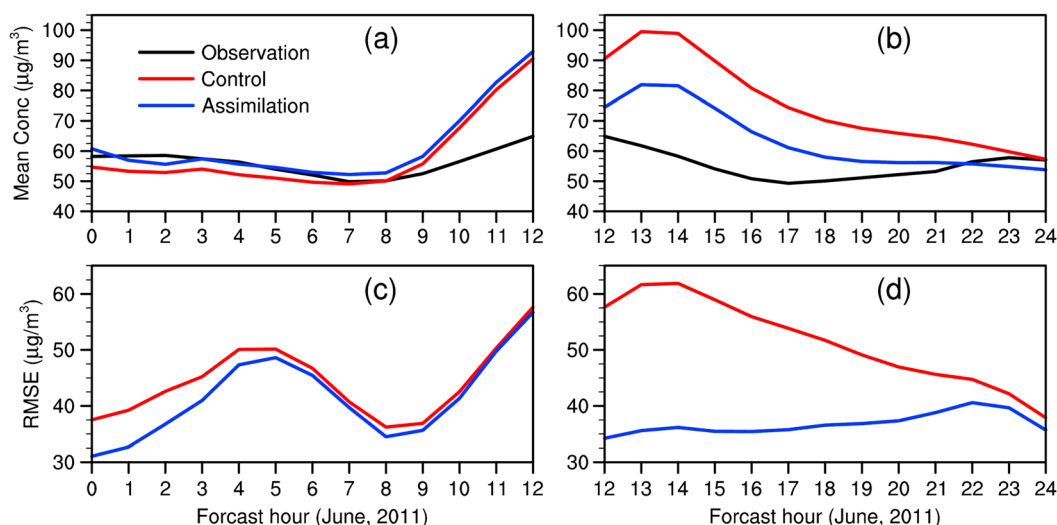


Figure 6. (top) Mean concentration ($\mu\text{g}/\text{m}^3$) and (bottom) root-mean-square error ($\mu\text{g}/\text{m}^3$) of surface PM_{2.5} forecasts as a function of forecast range, verified against observations from Nanjing and Shanghai during 1 to 28 June 2011. The forecasts are produced from the 0000 (left) and 1200 (right) UTC initializations.

the time-continuity check if it satisfies $|O(t) - O(t \pm 1)| \leq m(t)$, where the function $m(t)$ is determined empirically (e.g., $m(t) = 50 + 0.15O(t)$ for this study).

[19] As mentioned in section 2.3, the observation error covariance matrix $\mathbf{R}$ in equation (1) contains both measurement and representativeness errors. However, observation errors are not provided with observations, and no general formulation is available to estimate such errors. Similar to S12, the measurement error ε_o is defined as $\varepsilon_o = 1.5 + 0.0075 \cdot \Pi_o$, where Π_o denotes PM₁₀ observational values (units: $\mu\text{g}/\text{m}^3$). Thus, higher PM₁₀ values were associated with larger measurement errors. The representativeness error, also following Elbern *et al.* [2007] and S12, depended on the model resolution and the characteristics of the observation locations. The observation errors are assumed uncorrelated so that $\mathbf{R}$ is a diagonal matrix.

4. Experimental Design

[20] Two parallel experiments were performed to evaluate the impact of PM₁₀ DA on analyses and forecasts of aerosols over China. One experiment served as the control and did not employ any DA, while PM₁₀ DA was implemented in the other. Both experiments began from the same set of ICs valid at 0000 UTC 01 June 2011 that were spun-up over 5 days beginning 27 May, similar to Pagowski *et al.* [2010] and S12. Each experiment initialized a new WRF/Chem forecast every 6 h between 0000 UTC 01 June and 0000 UTC 28 June 2011. All 0000 and 1200 UTC initializations produced 12 h forecasts, while 6 h forecasts were produced from the 0600 and 1800 UTC initializations. Each initialization, both LBCs and ICs for meteorological fields were updated by interpolating GFS analyses onto the model domain, and gaseous chemical variables were initialized from the previous cycle's 6 h forecast.

[21] As described by L11 and S12, both experiments only differed regarding initialization of the 16 GOCART aerosol species contained in the GSI analysis. In the control

experiment, the initial GOCART aerosol fields were simply taken from the previous cycle's 6 h forecast. However, the PM₁₀ assimilation experiment implemented a 3DVAR aerosol analysis every 6 h, using the GOCART aerosol fields from the previous cycle's 6 h forecast as the background. The analyses were then used as aerosol ICs for subsequent WRF/Chem forecasts. Thus, the experiments only differed in that 3DVAR DA updated the GOCART aerosol species in one experiment but not the other. Both experiments used the same physical and chemistry options outlined in section 2.1.

[22] The analyses and forecasts from the two experiments were compared to surface PM₁₀ measurements. The results of these comparisons are now described.

5. Results

[23] This section presents results from the control and assimilation experiments outlined above. Since surface PM₁₀ DA had a very small impact on AOD (not shown), we only performed verification against MEP PM₁₀ data and surface PM_{2.5} observations from two sites in Nanjing and Shanghai. Although forecasts were produced every 6 h, we focus on examining the PM₁₀ DA impact on aerosol ICs at 0000 and 1200 UTC (local time = UTC + 8 h) before evaluating the subsequent WRF/Chem forecasts and then discuss the IPR differences in PM₁₀ formation. Here, three basic statistical measures, mean bias (BIAS), root-mean-square error (RMSE), and correlation coefficient (CORR), as described in Zhang *et al.* [2006], are applied to evaluate the results. When compared with surface PM₁₀ and PM_{2.5} observations, the model values at the lowest vertical level were horizontally interpolated to the observation sites.

5.1. Impacts on Aerosol ICs

[24] The model evaluation statistics (Table 1) from the control and assimilation experiment for the lowest model level PM₁₀ mass concentrations were calculated against observations from the MEP network over all 0000 and 1200

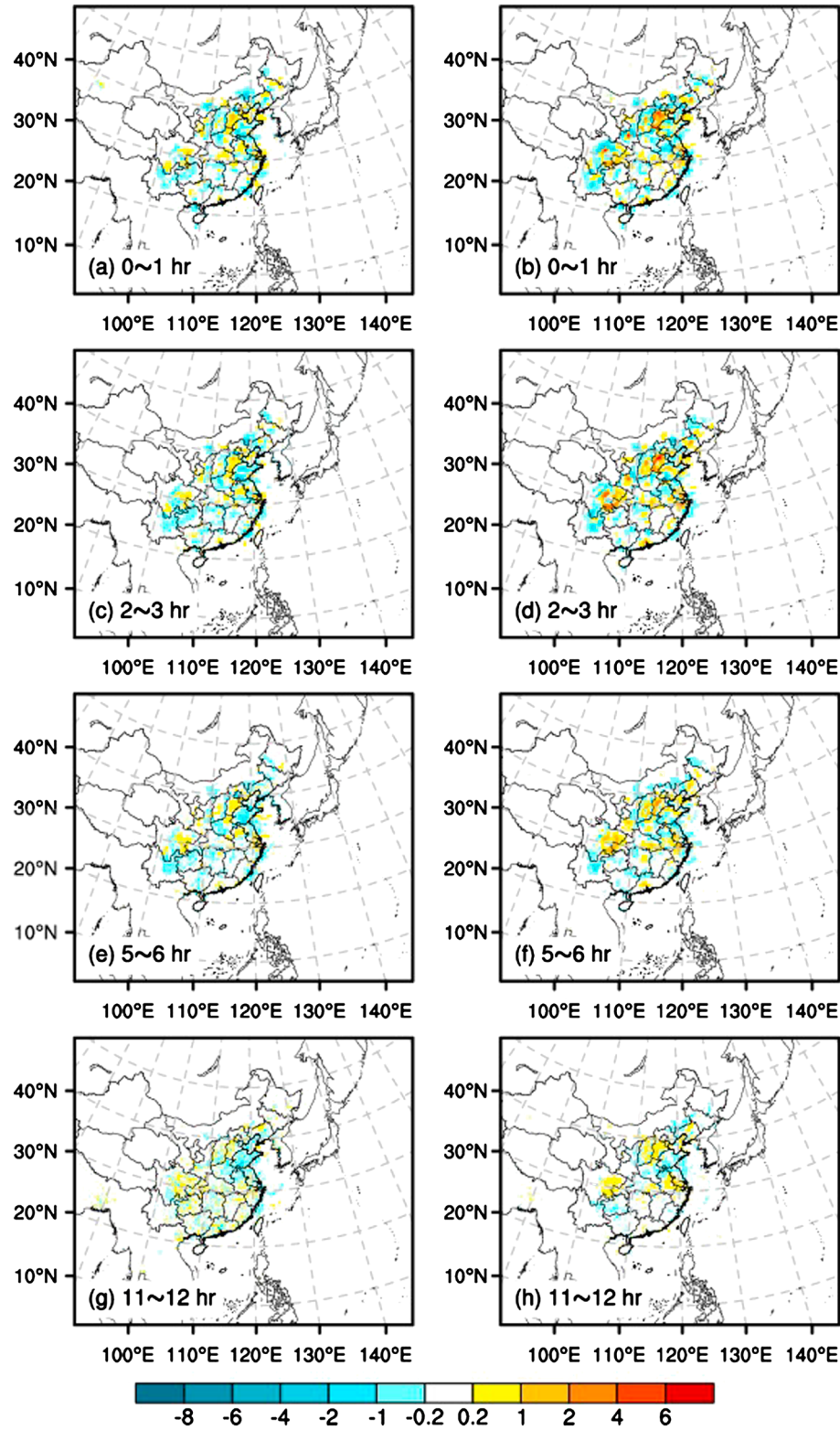


Figure 7. Spatial distribution of average differences of hourly PM₁₀ IPRs ($\mu\text{g}/\text{m}^3/\text{h}$) between the two experiments (assimilation minus control) advection processes (horizontal and vertical) at the lowest model level, aggregated over all 0000 (left) and 1200 (right) UTC initializations.

UTC initializations during 01 to 28 June 2011. There are significant systematic biases and large RMSEs for the control. The negative model bias of $-22.5 \mu\text{g}/\text{m}^3$ at 0000 UTC (morning) indicates a significant underestimation, about

29.1% lower than the observed PM₁₀ concentrations. Conversely, a positive bias of $14.8 \mu\text{g}/\text{m}^3$ at 1200 UTC (evening) indicates a significant overestimation, about 21.5% higher than the observed PM₁₀ values. Additionally, the

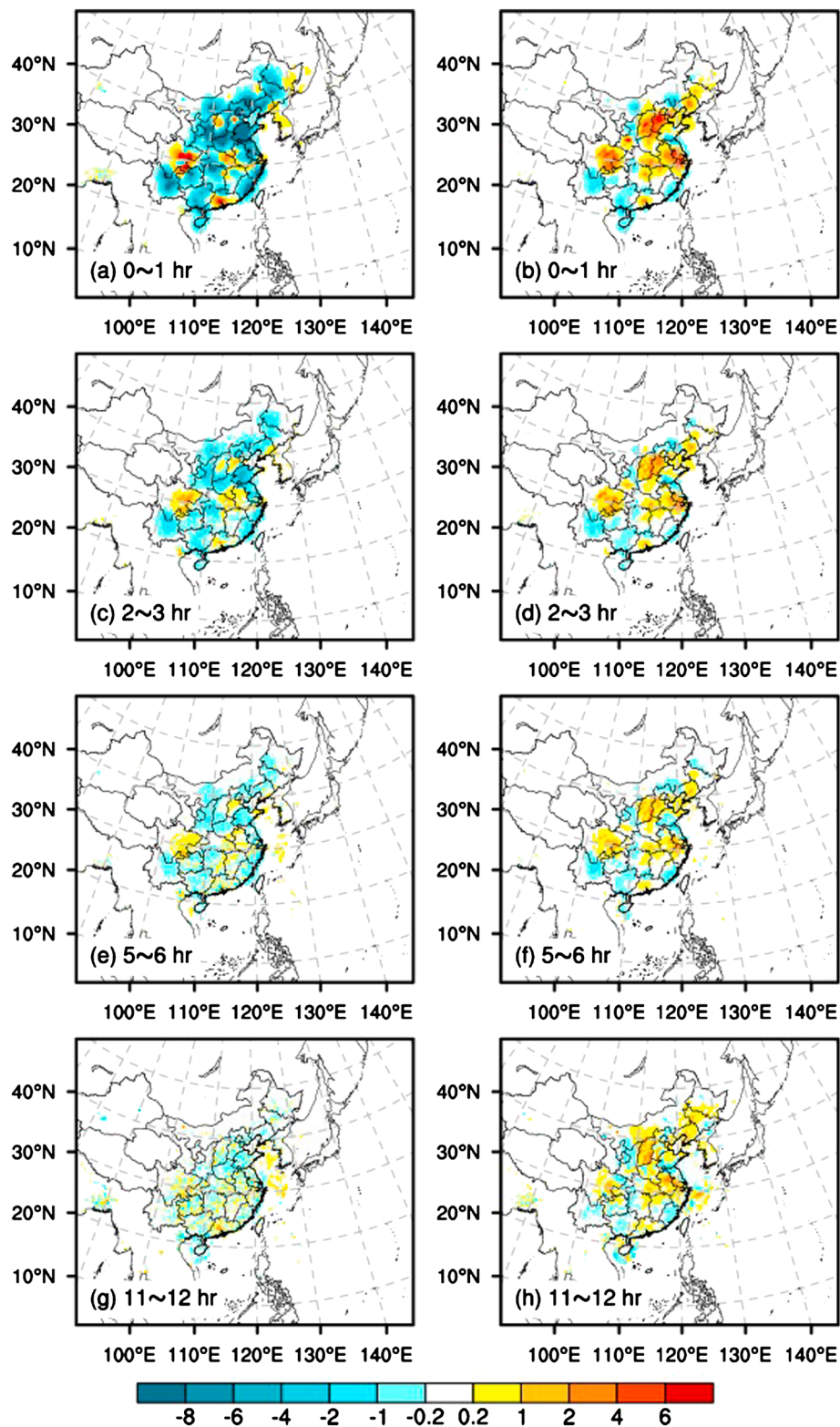


Figure 8. Same as Figure 7 except for vertical diffusion processes.

mean RMSE and CORR between the control and the observations is about $53.8 \mu\text{g}/\text{m}^3$ and 0.4, respectively. These magnitudes are similar to those of PM₁₀ forecast error statistics over the Yangtze River Delta region [Wang *et al.*, 2012] and the Pearl River Delta region [Chen *et al.*, 2009].

[25] After assimilating surface PM₁₀ observations, the statistics show much better agreement with observations than the control. Overall biases are dramatically reduced; RMSEs are decreased by about 60%, and CORRs are greatly increased from about 0.4 to 0.9. Scatter plots of simulated

PM₁₀ mass concentrations over all 0000 and 1200 UTC initializations from the control and assimilation experiments against observations are also illustrated in Figure 2. These results indicate that initial PM₁₀ fields can be adjusted effectively by our DA approach.

[26] Figure 3 depicts the spatial distribution of time-averaged PM₁₀ differences (assimilation minus control) at the lowest model level over all 0000 and 1200 UTC initializations. The patterns closely match the measurement sites in Figure 1. Consistent with Table 1, at 0000 UTC, the positive differences indicate PM₁₀ DA primarily increased the model PM₁₀, and the largest value (greater than 60 $\mu\text{g}/\text{m}^3$) occurs in Shandong province (Figure 3a). At 1200 UTC, the negative differences indicate PM₁₀ DA primarily decreased the model PM₁₀, and the lowest value (less than $-50 \mu\text{g}/\text{m}^3$) appears in South Hebei province (Figure 3b). The differences imply model errors were present, likely from inaccurate emission rates in time and space. For instance, some underestimation may come from crop residue burning due to summer harvest [Yang *et al.*, 2008], not properly represented in the emissions.

[27] The vertical profiles of simulated PM₁₀ mass concentrations from the control and assimilation experiments are given in Figure 4 separately for locations where the surface PM₁₀ differences were positive and negative. The extent of the vertical impact of PM₁₀ DA is mainly determined by the vertical correlation of BECs [Benedetti and Fisher, 2007]. Although the DA impact extended up to model level 18, ~ 3.2 km above ground level (AGL), the biggest impact was confined below model level 8 (~ 0.8 km AGL). The change of total aerosol mass due to PM₁₀ DA diminished rapidly as a function of height, which is consistent with S12.

[28] It is also interesting to examine how DA impacted the PM₁₀ vertical gradients below 1 km. Where aerosol mass was increased by PM₁₀ DA, the vertical gradient of difference was negative, indicating after DA vertical aerosol mass gradient became stronger (Figure 4a). Conversely, PM₁₀ DA produced far weaker mass gradients in the vertical where aerosol mass decrease occurred (Figure 4b). These results show that PM₁₀ DA greatly changed vertical structures of aerosol mass at near-surface model levels, which were also found by S12 when solely assimilating PM_{2.5}. Vertical gradient controls the vertical diffusion term in equation (1).

5.2. Forecast Verification

[29] To better understand the performance of the control and assimilation experiments, the BIAS, RMSE, and CORR of surface PM₁₀ forecasts as a function of forecast range were calculated against MEP observations separately for the 0000 and 1200 UTC initializations.

[30] The control experiment had biases (~ -30 to $20 \mu\text{g}/\text{m}^3$) that were characterized by the diurnal cycle (Figures 5a and 5b). The positive (negative) biases spanned 1100 (1700) to 1600 (1000) UTC, indicating that WRF/Chem overpredicted (underpredicted) the surface PM₁₀ in this time range. DA impact on systematic biases of aerosol forecasts depended greatly on sign of the initial and subsequent forecast biases from the control. Where the signs were same (opposite) between the initial and subsequent forecast biases, improvement (deterioration) on subsequent forecast biases was produced by PM₁₀ DA. Both forecasts from 0000 and 1200 UTC initializations had the same behaviors, which

were primarily driven by overall increase or decrease of model aerosol mass after DA.

[31] Even though the assimilation experiment had larger biases at times, PM₁₀ DA produced consistently lower RMSEs and higher correlations for the entire 12 h forecasts, reflecting the overall benefit of assimilating surface PM₁₀ observations (Figures 5c–5f). However, the differences between the experiments diminished with forecast lead time, likely due to model processes and emissions dominating at later periods [Kahnert, 2008]. Also, the forecast accuracy from the assimilation experiment decreased with time, as RMSEs increased and CORRs decreased. For instance, the relative RMSE reduction due to DA is 59.7% at 0000 UTC initialization and then decreases to 30.8%, 17.4%, 9.9%, and 0.5% for the first, third, sixth, and twelfth of forecasts, respectively (Figure 5c). Similarly, the relative correlation increase due to DA is 128.0% at 0000 UTC initialization and then decreases to 96.6%, 76.1%, 33.1%, and 11.5% for the corresponding forecast ranges (Figure 5e). Tombette *et al.* [2009] and Lee *et al.* [2013], who assimilated surface PM₁₀ observations over Europe and South Korea, respectively, also noticed a small DA impact beyond about 12 h forecast.

[32] In addition, we also evaluated PM_{2.5} forecasts at Nanjing and Shanghai (two megacities over Eastern China) for the whole experimental period. The model-simulated PM_{2.5} was diagnosed within WRF/Chem, as described in S12, and the surface PM_{2.5} mass concentration was measured using the TEOM method. Figure 6 shows statistics, as a function of forecast range, of mean PM_{2.5} mass concentrations from observation, control, and PM₁₀ DA experiments (top two panels) and RMSEs for the control and DA experiments verified against surface PM_{2.5} observations (bottom two panels). As in Figure 5, statistics were calculated separately for 0000 (left) and 1200 (right) UTC initializations. These results are overall consistent with Figure 5, even though the error magnitudes and variation with forecast range differed slightly. The rapidly decreasing DA impact is also evident in PM_{2.5} verification.

[33] This behavior of the short-lasting DA impact with the forecast lead time has also been found by Tombette *et al.* [2009] and Lee *et al.* [2013] when assimilating surface PM₁₀ observations and Pagowski *et al.* [2010] and S12 when assimilating surface PM_{2.5} data. S12 gave a qualitative discussion on this issue and concluded that advection and vertical mixing are the main causes for the lack of long-lasting surface observation impact. Likewise, Elbern *et al.* [2007] also remarked that optimizing the emission rate is far more important than improving ICs by assimilating surface ozone observations through the EnKF and 4DVAR approaches, indicating that the long-range forecast is much less sensitive to the initial state than the emissions. We implemented PA capability within the WRF/Chem model, which provides a method to quantify the contributions of various processes to PM₁₀ formation from both experiments, as detailed in the next subsection.

5.3. PA for PM₁₀ Formation

[34] As stated in section 2.2, while the IPR deals with the net effect of all the physical and chemistry processes on model simulation, the IPR differences between parallel experiments represent various process-contribution differences. The IPR diagnosed from the control experiment showed that

the emission process (vertical diffusion) was the predominant contributor to PM₁₀ production (loss) at the lowest model level. Advection (including horizontal and vertical) and dry deposition only played secondary roles in PM₁₀ formation, and the contribution from other processes (such as chemical reactions) was very small (not shown). These findings are consistent with *Chen et al.* [2009], *Zhang et al.* [2009b], and *Liu et al.* [2010], who examined the relative importance of major atmospheric processes in PM formation.

[35] In this study, the IPR differences, originated mainly from adjusted aerosol ICs due to PM₁₀ DA, were examined to identify the main cause of the short-lasting DA impact of surface observations. Figure 7 depicts the time-averaged spatial distribution of PM₁₀ IPR differences (assimilation minus control) of advection processes at the lowest model level for various forecast hours separately for 0000 and 1200 UTC initializations. The IPR differences have no significant change in the first three hours (up to $\pm 4 \mu\text{g}/\text{m}^3/\text{h}$) and become smaller beyond the fifth forecast hour. Differences were mainly caused by localized increment structures around the measurement sites (Figure 3). The PM₁₀ IPR differences of vertical diffusion process are also presented in Figure 8. The largest differences occur in the initial hour (up to $-20 \mu\text{g}/\text{m}^3/\text{h}$ on minimum and $8 \mu\text{g}/\text{m}^3/\text{h}$ on maximum) and then are reduced dramatically and rapidly with forecast hour, more intense at day (0000 to 1200 UTC) than at night (1200 to 0000 UTC). Note that negative values indicate more loss or less production of PM₁₀ in the assimilation experiment and vice versa. In particular, the negative (positive) differences for vertical mixing indicate the enhanced (reduced) upward diffusion from the ground or downward dilution from above.

[36] Compared to advection process, the differences of vertical mixing are larger in value and wider in area within the same time, dominating the IPR differences between the two experiments. Clearly, vertical diffusion was more important than advection, causing the rapidly diminishing DA impact. The IPR differences are relatively small for the other processes, and the time-averaged values are lower than $\pm 1.5 \mu\text{g}/\text{m}^3/\text{h}$ for all forecast ranges (not shown).

[37] Recall that the same GFS analyses were used in both experiments for meteorological ICs. Therefore, initial aerosol fields primarily determine the strength of advection and diffusion processes. We note that a good spatial coherence is shown between PM₁₀ mass differences and PM₁₀ IPR differences of vertical mixing at the lowest model level, where positive (negative) surface PM₁₀ differences mostly correspond to negative (positive) IPR differences (Figures 3 and 8). This further indicates that, where aerosol mass increase (decrease) occurred, enhanced (reduced) upward diffusion produces more dilution (accumulation) of surface aerosol mass. Thus, surface PM₁₀ differences weaken continuously, leading to rapidly diminishing DA impact with time, especially in the early forecast hours.

[38] As evidenced in Figure 4, vertical gradients of aerosol mass at initializations were markedly changed after assimilating surface PM₁₀ observations, also found by S12 when solely assimilating PM_{2.5}. Thus, whenever surface PM₁₀ observations are assimilated, sharper (weaker) vertical gradient due to DA enhanced (lessen) vertical mixing and quickly diluted (accumulated) surface aerosol mass, leading to the fast-fading DA impact especially in the early forecast hours.

Furthermore, *Tombette et al.* [2009], *Pagowski et al.* [2010], S12, and *Lee et al.* [2013] also noted that fast-fading DA impact was generated in the early forecast hours when assimilating surface PM observations. Conversely, L11 and S12 found that AOD DA refines the total aerosol mass throughout the column, maintaining similar vertical structures, which allows a longer-lasting DA impact. These results all underline the importance of a correct vertical profile of aerosol mass, implying some limitation of assimilating the surface observations into ICs.

6. Summary and Discussion

[39] The ability of assimilating hourly averaged surface PM₁₀ observations from the MEP network was added into the aerosol DA framework within the GSI 3DVAR analysis system. This system was applied to daily aerosol forecasts produced in June 2011 over China. In the assimilation experiment, an aerosol analysis was performed every 6 h to update the WRF/Chem aerosol variables. To evaluate the effectiveness of DA, a control experiment without DA was also performed. Results revealed that aerosol analyses matched PM₁₀ observations much better than the control, which improved the surface PM₁₀ forecasts up to at least 12 h in terms of RMSE and correlation. However, the forecast bias can be enlarged by DA when its sign is opposite to that at initial time. Overall, our findings suggest that model-driven aerosol biases [e.g., *McKeen et al.*, 2007; *Misenis and Zhang*, 2010; *Lin et al.*, 2008b; *Zhang et al.*, 2010] can be mitigated by aerosol DA. Further studies are needed to further document the processes leading to aerosol bias.

[40] As also found in earlier work, the DA impact diminished rapidly with forecast range when solely assimilating surface aerosol observations. PA was carried out within WRF/Chem to probe into the differences in PM₁₀ formation between the control and assimilation experiments. The results show that fast-fading DA impact on aerosol forecast, especially in the early forecast hours, mostly came from vertical mixing, with minor contributions from horizontal and vertical advections. This behavior is mainly related to unbalanced aerosol fields in the horizontal and vertical after assimilating surface observations into the initial model state. This implies the need for aerosol observations with vertical information for more accurate 1–2 day forecast of surface aerosols.

[41] In the current assimilation experiment, only the initial state of aerosol fields was adjusted by cyclic PM₁₀ DA, which limited the DA impact on long-range aerosol forecasts. There are some indications that emission sources appear to be the primary cause of the PM production [e.g., *Chen et al.*, 2009; *Zhang et al.*, 2009b; *Liu et al.*, 2010], implying that emission adjustments are critical to the model forcing needed to maintain aerosol profile adjustments due to the assimilation of surface PM₁₀ observations. As emission uncertainty is widely recognized as a major factor limiting the accuracy of aerosol forecasts, especially in areas with air quality problems [*Dubovik et al.*, 2008], it should be particularly promising to analyze sources of aerosols and their precursors. Recently, adjoint inverse modeling systems [e.g., *Hakami et al.*, 2005; *Elbern et al.*, 2007; *Yumimoto et al.*, 2007, 2008a; *Henze et al.*, 2009; *Ku and Park*, 2011] have allowed

a “top-down” emission optimization at the model grid scale, providing constraints on emissions from surface aerosol observations. It can be expected that emission optimization will produce better forecast results than simply adjusting the initial model state by assimilating surface observations [e.g., Henze *et al.*, 2009]. Furthermore, more desirable forecast improvements will likely be achieved by the joint adjustment of ICs and emission inventories. These issues will be addressed in future work.

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